



Nitrous Oxide Reduction by an Obligate Aerobic Bacterium, *Gemmatimonas aurantiaca* Strain T-27

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ABSTRACT N₂O-reducing organisms with nitrous oxide reductases (NosZ) are known as the only biological sink of N₂O in the environment. Among the most abundant *nosZ* genes found in the environment are *nosZ* genes affiliated with the understudied *Gemmatimonadetes* phylum. In this study, a unique regulatory mechanism of N₂O reduction in *Gemmatimonas aurantiaca* strain T-27, an isolate affiliated with the *Gemmatimonadetes* phylum, was examined. Strain T-27 was incubated with N₂O and/or O₂ as the electron acceptor. Significant N₂O reduction was observed only when O₂ was initially present. When batch cultures of strain T-27 were amended with O₂ and N₂O, N₂O reduction commenced after O₂ was depleted. In a long-term incubation with the addition of N₂O upon depletion, the N₂O reduction rate decreased over time and came to an eventual stop. Spiking of the culture with O₂ resulted in the resuscitation of N₂O reduction activity, supporting the hypothesis that N₂O reduction by strain T-27 required the transient presence of O₂. The highest level of *nosZ* transcription (8.97 *nosZ* transcripts/*recA* transcript) was observed immediately after O₂ depletion, and transcription decreased ~25-fold within 85 h, supporting the observed phenotype. The observed difference between responses of strain T-27 cultures amended with and without N₂O to O₂ starvation suggested that N₂O helped sustain the viability of strain T-27 during temporary anoxia, although N₂O reduction was not coupled to growth. The findings in this study suggest that obligate aerobic microorganisms with *nosZ* genes may utilize N₂O as a temporary surrogate for O₂ to survive periodic anoxia.

IMPORTANCE Emission of N₂O, a potent greenhouse gas and ozone depletion agent, from the soil environment is largely determined by microbial sources and sinks. N₂O reduction by organisms with N₂O reductases (NosZ) is the only known biological sink of N₂O at environmentally relevant concentrations (up to ~1,000 parts per million by volume [ppmv]). Although a large fraction of *nosZ* genes recovered from soil is affiliated with *nosZ* found in the genomes of the obligate aerobic phylum *Gemmatimonadetes*, N₂O reduction has not yet been confirmed in any of these organisms. This study demonstrates that N₂O is reduced by an obligate aerobic bacterium, *Gemmatimonas aurantiaca* strain T-27, and suggests a novel regulation mechanism for N₂O reduction in this organism, which may also be applicable to other obligate aerobic organisms possessing *nosZ* genes. We expect that these findings will significantly advance the understanding of N₂O dynamics in environments with frequent transitions between oxic and anoxic conditions.

KEYWORDS *Gemmatimonadetes*, nitrous oxide, nitrous oxide reduction, RT-qPCR

The recent meteorological data from the first half of the year 2016 witnessed the highest worldwide average temperature ever recorded for the same period of the year, supporting the concern of the scientific world about ongoing climate change

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(GISS Surface Temperature Analysis [<http://www.ncdc.noaa.gov/sotc/global/201606>]). The increase in greenhouse gas emission due to various anthropogenic activities has been regarded as the major culprit of the general upward trend in the global temperature and climate anomalies occurring with alarming frequency across the globe (1). Nitrous oxide (N_2O) is a greenhouse gas with a global warming potential ~300 times that of CO_2 and is the third most important contributor to global warming (6.2%) after CO_2 (78%) and CH_4 (16%) (1–4). Further, the abolition of chlorofluorocarbons has left N_2O as the largest contributor to the destruction of the ozone layer in the stratosphere (5, 6). Therefore, control of N_2O emissions is indispensable in the efforts to curb global warming and climate change.

Both natural and anthropogenic sources of N_2O have a predominantly biological origin, with biological transformation of N fertilizer applied to agricultural soils being the single largest source (1, 7, 8). Nitrification produces N_2O as a by-product of ammonia oxidation, and denitrification emits N_2O as a stable intermediate or an end product (8–11). Other relatively minor sources of N_2O include dissimilatory reduction to ammonium (DNRA) and chemodenitrification (8, 12–14). In contrast to the diverse pathways leading to the production of N_2O , the sole biological sink process of N_2O in the environment is its reduction by the organisms expressing nitrous oxide reductases (NosZ) (8, 15–17). N_2O reduction was originally regarded merely as a part of the denitrification cascade. N_2O reduction as an independent respiratory reaction had not garnered the deserved interest until recent discoveries unveiled the unexpectedly broad diversity of *nosZ* (15, 18). *nosZ* genes of the novel clade II are often found in organisms lacking the genes encoding the key denitrification enzymes, namely *nirK* or *nirS*, indicating that these organisms utilize N_2O reduction as a respiratory reaction independent from denitrification (19). Thus, these nondenitrifying N_2O reducers function as *de facto* sinks of N_2O , as confirmed by physiological observations in experiments with isolates possessing NosZ (20, 21). Indeed, a recent study on the kinetics of N_2O reductions revealed that the clade II *nosZ*-containing organisms have significantly higher affinity to N_2O than the clade I *nosZ*-containing organisms, supporting the hypothesis that the organisms with clade II *nosZ* may contribute to the mitigation of N_2O emissions from nonpoint sources (17).

Metagenomic analyses of environmental DNA have revealed that clade II *nosZ* genes are, in fact, abundant in diverse environments, ranging from tropical forest and hot desert to Arctic tundra and polar desert (18, 22). Among the most abundant phylogenetic groups of *nosZ* in the environment are clade II *nosZ* genes affiliated with the *Gemmatimonadetes* phylum (16, 18). *Gemmatimonadetes* have been identified as one of the most abundant phyla of bacteria in soil environments and often constitute >2% of the total bacterial population (23, 24); however, only three representative strains of this phylum (*Gemmatimonas aurantiaca* strain T-27, *Gemmatirosa kalamazoonesis* strain KBS708, and *Gemmatimonas phototrophica* strain AP64) have been isolated to date, and the physiological characteristics of these organisms are virtually unknown (25–28). Although clade II *nosZ* genes were found in the genomes of *G. aurantiaca* strain T-27 and *G. kalamazoonesis* strain KBS708, these strains were both characterized as obligate aerobes, and respiration on any other electron acceptors, e.g., N_2O , has yet to be explored (25–27). Therefore, questions remain unanswered regarding the functionality of this *Gemmatimonadetes* NosZ and its potential physiological role as a respiratory enzyme. As this particular group of *nosZ* genes constituted up to 33% of the entire *nosZ* gene pools identified in soil metagenomes, an understanding of the N_2O reduction phenotype of the *Gemmatimonadetes* phylum is crucial for predicting N_2O sink capabilities of subsurface soil environments. In this study, N_2O reduction by *G. aurantiaca* strain T-27 was observed in both the absence and presence of oxygen. The inability of this organism to consume N_2O in the complete absence of oxygen and the unexpected transcription pattern of *nosZ*, i.e., upregulation in the presence of oxygen and downregulation in the absence of oxygen, suggest a novel regulatory mechanism for N_2O respiration by obligate aerobic microorganisms.

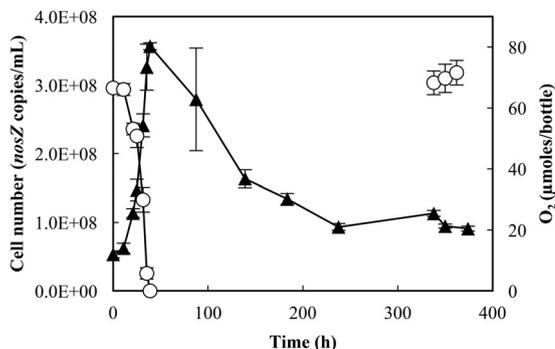


FIG 1 Aerobic growth of *G. aurantiaca* strain T-27 cultures in partially oxic conditions. The changes in amounts of O₂ (○) and *nosZ* copy numbers (▲) in the reaction bottles were monitored. Additional O₂ was added at $t = 339$ h. The data points are averages of duplicate experiments, with the error bars representing the standard deviations.

RESULTS

Aerobic growth of *G. aurantiaca* strain T-27. *Gemmatimonas aurantiaca* strain T-27 was incubated with $66.6 \pm 1.1 \mu\text{mol O}_2$ as the electron acceptor to obtain its growth curve under aerobic conditions (Fig. 1). The cell number, quantified by determining the *nosZ* gene copy number, increased from $5.5 \times 10^7 \pm 0.5 \times 10^7 \text{ cells} \cdot \text{ml}^{-1}$ to $3.6 \times 10^8 \pm 0.5 \times 10^8 \text{ cells} \cdot \text{ml}^{-1}$ at an exponential-growth rate of $0.068 \pm 0.004 \cdot \text{h}^{-1}$ (see Fig. S1 in the supplemental material). The exponential-growth phase lasted until O₂ was depleted. Cell decay began immediately after O₂ depletion; there was no stationary phase. The cell number decreased by more than 3-fold, from $3.6 \times 10^8 \pm 0.5 \times 10^8 \text{ cells} \cdot \text{ml}^{-1}$ to $1.1 \times 10^8 \pm 0.4 \times 10^8 \text{ cells} \cdot \text{ml}^{-1}$, over 299 h of anoxic incubation following O₂ depletion. As $66.6 \pm 1.1 \mu\text{mol O}_2$ was consumed, the cellular yield on O₂ respiration was calculated to be $1.14 \times 10^8 \pm 0.03 \times 10^8 \text{ cells} (\mu\text{mol e}^- \text{eq})^{-1}$. The addition of O₂ at $t = 338$ h failed to revitalize the inactivated *G. aurantiaca* strain T-27 cells, as three consecutive measurements at 12 h-intervals confirmed no significant increase in cell number or decrease in the amount of O₂. Unexpectedly, no significant decrease in glucose concentration was observed throughout the experiment (data not shown), indicating that other organic compounds in the complex medium NM-1 (polypeptone, monosodium glutamate, and/or yeast extract) also served as electron donors and sources of carbon for *G. aurantiaca* strain T-27.

N₂O consumption under anoxic and partially oxic conditions. Initially, *G. aurantiaca* strain T-27 was examined for N₂O reduction activity under completely anoxic conditions (Fig. 2A). The amount of N₂O in the culture bottle after 312 h of incubation ($105.1 \pm 6.4 \mu\text{mol/bottle}$) was not significantly different ($P > 0.05$) from the initial amount of N₂O ($103.5 \pm 1.7 \mu\text{mol/bottle}$). When *G. aurantiaca* strain T-27 cells were incubated with O₂ added to the headspace ($55.8 \pm 3.9 \mu\text{mol/bottle}$) along with N₂O ($125.8 \pm 9.9 \mu\text{mol/bottle}$), active N₂O reduction was observed, but only after O₂ was depleted (Fig. 2B). The O₂ concentration dropped below the detection limit of the oxygen meter (0.02% in the headspace) within 25 h of inoculation, and the consumption of N₂O started upon the depletion of O₂. N₂O was depleted after 138.5 h of anoxic incubation. The observations from these anoxic and partially oxic incubations of strain T-27 with N₂O suggested that N₂O reduction activity in strain T-27 is dependent on O₂. No decrease in N₂O concentration was observed in killed-cell controls over 336 h of incubation (Fig. S2), confirming that the N₂O consumption was entirely biological.

Long-term N₂O consumption in the absence of oxygen. In a subsequent experiment under identical experimental conditions, N₂O consumption by *G. aurantiaca* strain T-27 was monitored over the long term (392 h) to investigate whether N₂O reduction activity is sustained in the absence of O₂ (Fig. 3A). Neither the cell yield ($1.27 \times 10^8 \pm 0.18 \times 10^8 \text{ cells} [\mu\text{mol e}^- \text{eq}]^{-1}$) nor the exponential-growth rate ($0.081 \pm 0.009 \text{ h}^{-1}$) during the initial aerobic growth was significantly different from

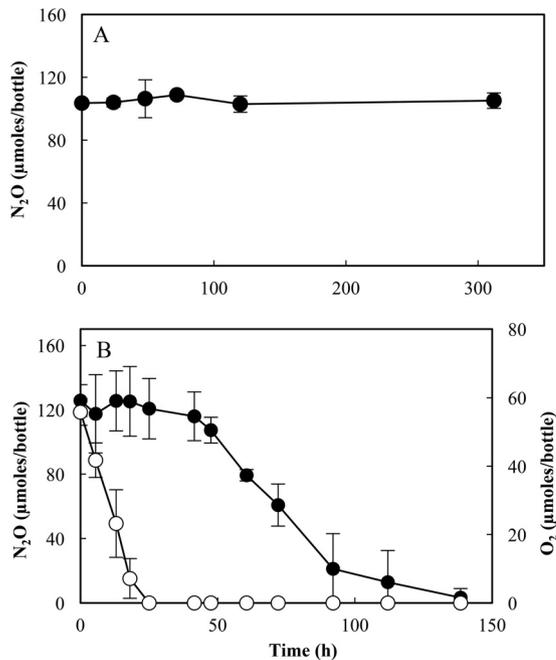


FIG 2 N₂O (●) and O₂ (○) consumption by *G. aurantiaca* strain T-27 cultures in the absence of O₂ (A) and initially amended with $55.8 \pm 3.9 \mu\text{mol O}_2$ (B). The data points are averages of triplicate experiments, with the error bars representing the standard deviations of the triplicate measurements.

the value determined from the aerobic incubation without N₂O (Fig. S1). The initial batch of N₂O ($125.8 \pm 0.9 \mu\text{mol}$) was rapidly consumed within 81.5 h after O₂ depletion. After an injection of an additional $133.1 \pm 2.7 \mu\text{mol N}_2\text{O}$, strain T-27 continued to consume N₂O until the reaction slowed down significantly after $t = 237.5$ h and came to a near-complete stop at $t = 338$ h, after a total of $159.9 \mu\text{mol N}_2\text{O}$ was consumed. N₂O respiration did not appear to be coupled to growth in strain T-27, as no significant increase in cell number ($P > 0.05$) was observed after the depletion of O₂. Unlike when strain T-27 was incubated without N₂O, O₂ depletion did not lead to an immediate decrease in cell number. The cell number was sustained for 144.5 h after the depletion of O₂, although a statistically insignificant ($P > 0.05$) drop from $3.6 \times 10^8 \pm 0.3 \times 10^8$

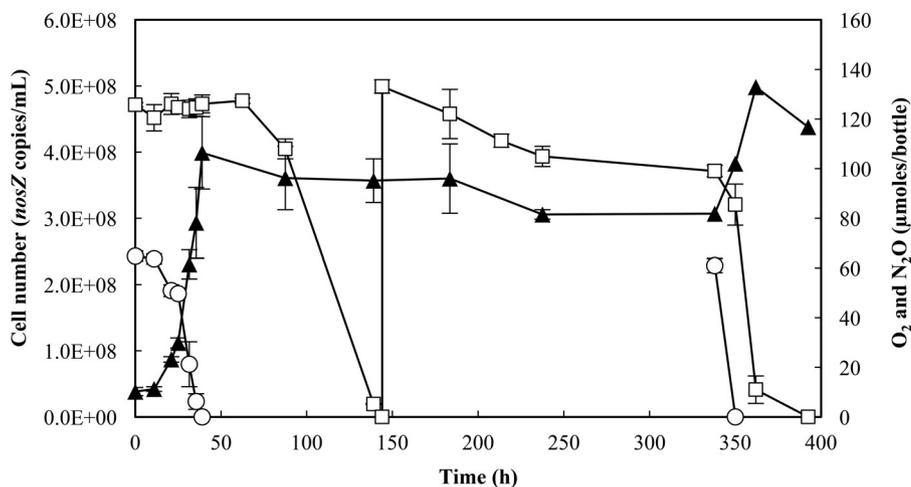


FIG 3 Long-term incubation of *G. aurantiaca* strain T-27 with N₂O after initial growth on O₂. The amounts of N₂O (□) and O₂ (○) in the reaction bottles were monitored, and the copy numbers of *nosZ* genes (▲) were measured at selected time points using qPCR for cell counts. N₂O was added upon N₂O depletion at $t = 144$ h, and O₂ was added at $t = 338$ h. The data points are averages of triplicate experiments, with the error bars representing the standard deviations of the triplicate measurements.

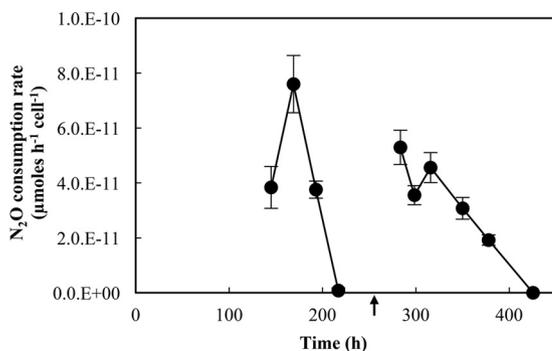


FIG 4 Monitoring of the change in rates of N₂O consumption by *G. aurantiaca* strain T-27. Each data point represents the N₂O consumption rate determined from linear regression of time versus N₂O amount (the average of triplicate samples) data. The error bars represent the standard errors of linear regression analyses. The arrow mark indicates the time of O₂ addition.

copies · ml⁻¹ to $3.1 \times 10^8 \pm 0.8 \times 10^8$ copies · ml⁻¹ was observed after that time point. The addition of O₂ at $t = 338$ h resulted in immediate consumption of O₂ accompanied by an immediate increase in the *nosZ* gene copy number to $5.0 \times 10^8 \pm 0.7 \times 10^8$ copies · ml⁻¹, followed by resuscitation of N₂O consumption. Residual N₂O (99.0 ± 2.3 mol) at the time of O₂ injection was rapidly reduced within 42.5 h of O₂ depletion. The different responses of the cultures amended with and without N₂O to O₂ depletion suggested that N₂O reduction helped sustain cell integrity and, in part, metabolic activities during transient anoxia.

In another set of experiments, the rates of N₂O consumption were measured at different time points during anoxic incubation following initial growth on O₂ (Fig. 4). The N₂O consumption rate peaked with $7.6 \times 10^{-11} \pm 1.0 \times 10^{-11}$ μmol · h⁻¹ · cell⁻¹ at $t = 169$ h (132 h after O₂ depletion) and decreased to zero at $t = 217$ h. N₂O reduction resumed after consumption of added O₂, and the maximum N₂O consumption rate, $5.3 \times 10^{-11} \pm 0.6 \times 10^{-11}$ μmol · h⁻¹ · cell⁻¹, was measured immediately after O₂ depletion. The maximum N₂O consumption rate was 69.7% of the highest value measured before resuscitation; however, N₂O consumption was sustained for a longer period of time after (144 h) than before (95 h) resuscitation. These kinetics measurements confirmed that N₂O consumption is not sustainable in the prolonged absence of O₂.

***nosZ* transcription analyses.** The transcription of *nosZ* was monitored with the samples withdrawn from a replicate of the batch culture used for monitoring of N₂O consumption and cell growth (Fig. 5, Fig. S3). The transcription of *nosZ* increased significantly, from 0.21 ± 0.21 *nosZ* transcripts/*recA* transcript upon inoculation to 8.97 ± 0.90 *nosZ* transcripts/*recA* transcript, immediately after the consumption of 63.1 μmol O₂ ($P < 0.05$). The transcription of *nosZ* decreased to 0.15 ± 0.23 *nosZ* transcripts/*recA* transcript 419 h after O₂ depletion, and the downregulation of *nosZ* transcription explained the disappearance of N₂O reductase activity. As expected from the resumption of N₂O reduction activity after injection of O₂, O₂ addition resuscitated *nosZ* transcription, and the maximum transcription level (1.86 ± 0.16 *nosZ* transcripts/*recA* transcript) was observed 91 h after O₂ depletion. This elevated transcription activity was sustained for >300 h after the oxygen depletion, even though a notable decreasing trend was observed after the peak, suggesting a less stringent regulation of *nosZ* transcription than that observed after the first O₂ depletion event. This observation was consistent with a less precipitous decrease in N₂O reduction rate after resuscitation than before O₂ addition. The highest peaks in *nosZ* transcription were observed immediately after O₂ was depleted, indicating that *nosZ* transcription occurred in the presence of O₂. These observations suggested that *G. aurantiaca* strain T-27 may require N₂O reductases expressed in the presence of O₂ to utilize N₂O as a surrogate electron acceptor in the anoxia that follows.

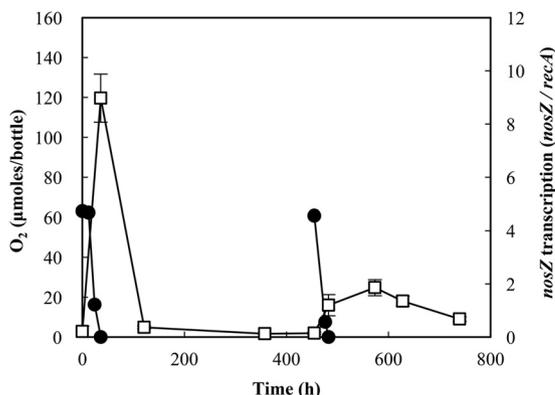


FIG 5 *nosZ* transcription in the *G. aurantiaca* strain T-27 culture during the long-term anoxic incubation with N₂O following initial partial-oxic incubation. The samples for the RT-qPCR analysis were collected at 9 crucial time points from the same culture vessel examined in Fig. S3. *nosZ* transcription level (□) was presented as the *nosZ* transcript copy numbers normalized with *recA* transcript copy numbers. O₂ (●) concentration is shown to indicate the time points of O₂ addition and depletion. The *nosZ* transcription data points are averages of triplicate samples treated independently through RNA extraction, purification, and reverse transcription procedures, and the error bars represent their standard deviations.

DISCUSSION

The *Gemmatimonadetes* phylum is one of the most abundant phylogenetic groups of microorganisms that constitute the soil bacterial communities (29, 30). Apart from the synthesis of a novel carotenoid by *G. aurantiaca* strain T-27 (31), not much has been reported regarding their physiology until recently, mainly due to the paucity of available isolates. The recent discovery of a full operon encoding type 2 photosynthetic reaction centers in the *G. phototrophica* strain AP-64 genome suggested that the *Gemmatimonadetes* phylum of bacteria may have adopted an oxygen-independent energy-generating metabolism to supplement an oxygen-dependent heterotrophic lifestyle in cases of transient hypoxia or anoxia (32). The utilization of N₂O by *G. aurantiaca* strain T-27 observed in this study may be interpreted in a similar context. Strain T-27 was previously characterized as an obligate aerobic bacteria unable to grow with any nonoxygen electron acceptors (26), and our experiments have confirmed that the presence of N₂O alone could not initiate N₂O reduction under conditions of complete anoxia or support growth; however, in the short term, N₂O utilization as the surrogate electron acceptor helped sustain cell viability in the absence of O₂. Similar transient enhancement of reduction of nonoxygen electron acceptors, i.e., NO₂⁻ and NO, was previously observed in obligate aerobic nitrifiers under hypoxic or semioxic conditions (33, 34). Geets et al. (34) proposed the hypothesis that one of the possible physiological roles of this nitrifier denitrification may be related to survival during anoxic periods, but this hypothesis has not yet been supported by experimental evidence. In recent research with *Nitrosomonas eutropha* strain C91, NO₂ gas was found to have a similar role of sustaining cell viability in oxygen-depleted cultures (35). The transcriptomics and proteomics data in the study suggested that *N. eutropha* strain C91 tuned down the assimilatory metabolism while retaining or stimulating dissimilatory metabolism upon anoxic incubation with NO₂ as the surrogate oxidant. A Gram-positive obligate aerobic bacterium, *Streptomyces coelicolor* A3(2), was also found to utilize its Nar-type respiratory nitrate reductase to survive extended periods of anoxia, although no growth was observed in the absence of O₂ (36, 37). The utilization of N₂O as a temporary substitute for O₂ in *G. aurantiaca* strain T-27 may be another such lifestyle that strictly aerobic microorganisms have adopted to generate cell maintenance energy to survive through periods of anoxia. In fact, in support of this hypothesis, several other organisms that have been characterized as obligate aerobes harbor clade II *nosZ* genes, including *Leptospira* spp., *Runella slithyiformis*, *Haliscomenobacter hydrossis*, and *Marivirga tractuosa* (38–41).

The physiology of N₂O reduction by *G. aurantiaca* strain T-27 observed in this study

was distinguishable from that of previously studied denitrifiers or nondenitrifying N_2O reducers in that (i) N_2O reduction was decoupled from cell growth and (ii) N_2O reduction activity was dependent on the transient presence of O_2 and dissipated under prolonged anoxia. Although N_2O reductase is a soluble periplasmic enzyme, earlier studies have demonstrated that the electron transfer chain to N_2O reductase involves cytochromes and thus is coupled to proton translocation across the membrane and energy conservation (42, 43). In fact, diverse groups of denitrifiers and nondenitrifying N_2O reducers were able to grow with N_2O as the sole electron acceptor (17, 43, 44). Albeit with different efficiencies and concentration thresholds, organisms with active clade I and clade II N_2O reductases examined in a recent kinetics study invariably coupled N_2O reduction to growth, and N_2O reduction entirely decoupled from cell growth has not been previously observed.

A plausible explanation for nongrowth in the absence of O_2 may be that O_2 is required for synthesis of metabolites necessary for strain T-27 to reproduce while N_2O can be used only as a temporary substitute for oxygen, as it can provide energy but not these necessary metabolites. A similar phenomenon was previously observed with *Campylobacter jejuni* strain 11168, a gastrointestinal pathogen capable of utilizing NO_3^- , NO_2^- , fumarate, and trimethylamine-*N*-oxide as electron acceptors for energy conservation under microaerophilic conditions, but not in the complete absence of oxygen (45). The oxygen dependence of *C. jejuni* strain 11168 was attributed to the requirement for oxygen-dependent ribonucleotide reductase (RNR). Another plausible explanation for the oxygen requirement of strictly aerobic bacteria is the use of HemF, the oxygen-dependent coproporphyrinogen III oxidase, for coproporphyrinogen III decarboxylation in heme synthesis (46). Neither case applies to *G. aurantiaca* strain T-27, as it possesses the genes encoding class II RNR and HemN, the oxygen-independent counterparts for class I RNR and HemF, respectively, in its genome; however, it is still possible that other essential metabolic functions for biosynthesis in *G. aurantiaca* strain T-27 may be carried out by oxygen-dependent enzymes.

Nitrous oxide reductase, like other enzymes in the denitrification cascade, has long been known to be sensitive to O_2 concentration (47, 48). That *G. aurantiaca* required the transient presence of O_2 for expression and activation of nitrous oxide reductases is nothing new. *Ensifer meliloti* strain 1021 was found to require partial oxygenation before activation of anaerobic denitrification and N_2O reduction, and *Paracoccus denitrificans* exhibited the highest level of *nosZ* expression immediately following O_2 depletion (49, 50). Nevertheless, the eventual termination of N_2O reduction in prolonged anoxia has not been observed in these organisms. In *E. meliloti* strain 1021, the expression levels of denitrification-related genes *napA*, *nirS*, *nosC*, and *nosZ* were sustained at an order-of-magnitude-higher level under completely anoxic conditions than under partially oxic conditions, suggesting continued expression and activity of denitrification and N_2O reduction (51). In *P. denitrificans*, the high level of *nosZ* transcription was sustained in the anoxic phase until all nitrogen oxides were depleted, indicating that NosZ enzymes synthesized during anoxic incubation were utilized for N_2O reduction (52). Further, in four strains of bacteria examined for kinetic properties, N_2O served as the sole source of electron acceptor for exponential growth in the complete absence of O_2 (17). As nitrous oxide reductase is active only in the absence of O_2 and thus is supposedly more beneficial for its owners under anoxic conditions (52), the disappearance of *nosZ* transcription activity in oxygen-depleted cultures of *G. aurantiaca* strain T-27 is rather unexpected and is unprecedented in any studies on denitrifiers or nondenitrifying N_2O reducers alike. These novel findings substantially expand the knowledge of microbial processes that contribute to N_2O emission mitigation, as this study demonstrated a novel regulatory mechanism that may be more generally applicable to *nosZ*-harboring obligate aerobic bacteria, which constitute an appreciable portion of the *nosZ*-harboring population in the environment (15, 18).

Subsurface soil environments alternate between oxic and anoxic conditions due to both natural and anthropogenic events. Precipitation events often lead to flooding of the vadose zone, hindering mass transfer of atmospheric O_2 into the subsurface soil

environment (53). In fertilized grassland, soil compaction from frequent mowing may also result in temporarily diminished O₂ availability (54). Elevated N₂O emissions have been observed during such transitions from oxic to anoxic conditions (55–57). The obligate aerobic organisms harboring *nosZ* genes, including *G. aurantiaca* strain T-27, may benefit from the elevated local N₂O concentrations, as they are capable of utilizing N₂O as a temporary surrogate for O₂ to survive temporary anoxia. These aerobic N₂O reducers may be undertaking an unexpectedly important role in reducing the amounts of N₂O emitted to the atmosphere in events of oxic-to-anoxic transitions. Future research is warranted for the development of novel experimental methods for an investigation of the true contribution of these obligate aerobic organisms to N₂O emission reduction, as enrichment-based investigations would not be able to capture these important N₂O sinks.

MATERIALS AND METHODS

Bacterial culture and growth conditions. *Gemmatimonas aurantiaca* strain T-27 was acquired from the Japan Collection of Microorganisms (JCM 11422). The culture medium used in this study was developed from NM-1 medium, a semidefined medium previously developed for isolation and culturing of *G. aurantiaca* strain T-27 (26). The medium contained, per liter, 0.5 g of glucose, 0.5 g of polypeptone, 0.5 g of monosodium glutamate, 0.5 g of yeast extract, 0.44 g of K₂HPO₄, 0.1 g of (NH₄)₂SO₄, and 0.1 g of MgSO₄. As nitrous oxide reductase is a copper-dependent enzyme and a lack of copper may hinder the activity of the enzyme (58), CuCl₂ was added to a concentration of 5 μM. The pH was adjusted to 7.0 with 5 N NaOH solution. For preparation of anoxic or partially oxic cultures, 100-ml aliquots of the medium were distributed into 160-ml serum bottles (Wheaton, Millville, NJ) and flushed with >99.999% N₂ for ~20 min to remove O₂. This degassing step was omitted for oxic precultures, prepared with 50 ml of medium in 160-ml serum bottles. The serum bottles were then sealed with butyl rubber stoppers (Geo-Microbial Technologies, Ochelata, OK). After autoclaving, 10% of headspace (6 ml of 60-ml headspaces) was aseptically replaced with air to prepare partially oxic cultures (~2.1% O₂ concentration in the headspace). No air was added to anoxic controls. Half a milliliter of 200× Wolin's vitamin solution (59) was added, and 2.0 ml of *G. aurantiaca* strain T-27 preculture was inoculated. Three milliliters of >99.999% N₂O (Deokyang Co., Ulsan, South Korea) was aseptically added using a disposable syringe connected to 0.2-μm-pore-size syringe filters (Advantec, Inc., Tokyo, Japan) after the same volume of headspace gas was removed. All culture vessels were incubated at 30°C with shaking at 140 rpm. Glucose was added in excess (0.5 g/liter), as a theoretical stoichiometric calculation estimated that only 15.0% (42.5 μmol) of added glucose would be consumed to reduce 125.8 μmol O₂ and 258.9 μmol N₂O (the maximum amounts of electron acceptors used in this research), even if glucose is used as the sole electron donor. Glucose concentrations were measured after the reactions were completed to confirm that the amount of electron donor added was not a limiting factor for the growth of *G. aurantiaca* strain T-27.

Analytical procedure. The amounts of N₂O in the serum bottles were quantified using an HP6890 series gas chromatograph equipped with an HP-PLOT/Q column and an electron capture detector (Agilent Technologies, Santa Clara, CA). The injector, oven, and detector temperatures were set to 200, 85, and 250°C, respectively (17). For each measurement, 200 μl of headspace gas was removed using a 1700 series gas-tight syringe (Hamilton Company, Reno, NV), and 100 μl of the withdrawn sample was manually injected into the gas chromatograph. The syringe was flushed at least three times with pressurized N₂ gas to remove O₂ before use, and 200 μl of N₂ was added upon each sampling event to prevent a pressure drop in the culture bottles. The change in the amounts of N₂O due to gas sampling was accounted for in the subsequent calculations (60). Oxygen concentrations were monitored with fiber-optic oxygen sensor spots and a FireStingO2 oxygen meter (Pyroscience, Aachen, Germany). The total amounts of N₂O and O₂ in the reaction vessels were calculated from the headspace concentrations as described previously (17). The dimensionless Henry's constants (calculated as moles in the headspace/ moles in the aqueous phase) of N₂O and O₂ at 30°C were calculated to be 1.92 and 33.3, respectively (61). These dimensionless Henry's constants were used to calculate the aqueous concentrations of N₂O and O₂. The amounts of the gases in the headspace and the aqueous phase were summed to determine the total amounts of N₂O and O₂ in the vessels. Glucose concentrations were measured colorimetrically using a glucose (HK) assay kit (Sigma-Aldrich, St. Louis, MO).

Monitoring of N₂O consumption. A series of incubation experiments was performed to investigate the consumption of N₂O by *G. aurantiaca* strain T-27. Oxic precultures for the experiments were prepared from glycerol stocks of strain T-27. For anoxic control experiments, 50 ml of the oxic precultures was harvested at the late-exponential phase (optical density at 600 nm [OD₆₀₀], ~0.043) and centrifuged at 4,800 × g for 10 min at 4°C. In an anaerobic chamber (Coy Laboratory Products, Inc., Grass Lake, MI) filled with 95% N₂ and 5% H₂, the pellets were equilibrated for 1 h and resuspended into 2 ml of medium taken from the anoxic culture bottles prepared as described above. The concentrated precultures were re-injected into the anoxic serum bottle, and the N₂O concentration was monitored for 312 h. After confirmation of the absence of N₂O consumption activity in the anoxic cultures, the N₂O consumption experiments were performed in partially oxic culture bottles prepared with a low concentration (2.1%) of O₂ in the headspace. One milliliter of frozen stock of strain T-27 was grown in 50 ml of oxic medium in a 160-ml serum bottle until the late-exponential phase. Each partially oxic culture bottle (with 100 ml

of medium) was inoculated with 2 ml of these late-exponential-phase cultures. O₂ concentrations were monitored until the concentrations dropped below the detection limit of the oxygen meter, and N₂O concentrations were measured at 5- to 26.5-h intervals until no remaining N₂O was detected. The killed-cell negative controls were inoculated with autoclaved aerobic precultures harvested at the late-exponential phase and concentrated by centrifugation. The concentrations of O₂ and N₂O were monitored for 336 h at 30°C.

Long-term incubation of *G. aurantiaca* strain T-27 on N₂O. In order to examine whether O₂ is required for sustenance of N₂O reduction activity, *G. aurantiaca* strain T-27 culture was initiated with 51.5 μmol O₂ and 122.6 μmol N₂O. Each time N₂O was depleted, 3 ml (nominally 121.6 μmol) of N₂O was aseptically injected into the culture bottle. After N₂O consumption came close to a halt, 6 ml of headspace was aseptically replaced with air, providing an additional 51.5 μmol O₂, to examine whether O₂ addition resuscitates N₂O reduction activity. O₂ and N₂O concentrations were monitored throughout the experiment at appropriate time intervals (3.5 to 100.5 h), and aqueous-phase samples were collected for quantification of *nosZ* genes and/or transcripts. As strain T-27 contains a single *nosZ* copy in its published genome (accession no. [NC_012489.1](#)), the copy number of the *nosZ* gene was used as a surrogate for cell counts. In the no-N₂O controls amended only with 51.5 μmol O₂, O₂ concentrations and cell numbers were monitored until 323 h after O₂ depletion. After O₂ replenishment, three measurements of O₂ concentrations and *nosZ* copy numbers were made at 12-h intervals.

In an independent set of experiments, the decrease in N₂O consumption rate was monitored in anoxia following initial O₂ consumption (51.5 μmol). Three independent cultures of *G. aurantiaca* strain T-27 were prepared identically to the above-described experiments. After O₂ depletion and onset of N₂O consumption, N₂O consumption rates were measured by taking at least three consecutive measurements of N₂O concentrations at a constant time interval (intervals were varied depending on the rates of N₂O consumption) and performing linear regression to calculate the slope of the time-versus-amount curves for the N₂O consumption rates. The average of the time points was taken as the representative time for each rate measurement. After each rate determination, N₂O in the culture bottles was replenished to the initial amount (~121.6 μmol N₂O/bottle) by adding >99.999% N₂O aseptically through 0.2-μm-pore-size filters. One milliliter of the aqueous phase was extracted immediately after N₂O replenishment for cell counting by quantitative PCR (qPCR), and the volumetric loss was replaced with fresh anoxic NM-1 medium. The N₂O consumption data were normalized, with the numbers of *nosZ* genes measured by qPCR. Upon cessation of N₂O reduction, 51.5 μmol O₂ was added, and N₂O reduction rates were monitored using the same protocol.

Cell culture sampling for DNA/RNA extraction. A qPCR technique was used for quantification of cell numbers, and reverse transcription-quantitative PCR (RT-qPCR) analyses were performed to examine the effect of O₂ on transcription of *nosZ*. A cell suspension was sampled at the time points determined to be crucial for observation of N₂O reduction by *G. aurantiaca* strain T-27. Each sampling was performed by withdrawing 1.0 ml (samples subjected to either qPCR or RT-qPCR) or 1.6 ml (samples subjected to both qPCR and RT-qPCR) of the aqueous phase using a sterile disposable syringe flushed with N₂ gas. To avoid pressure loss in the vessel, the same volume of N₂ gas was added upon each sampling event. Triplicate 0.2-ml aliquots were transferred to 1.5-ml DNase- and RNase-free tubes (Eppendorf, Hamburg, Germany) for DNA extraction. After centrifugation at 15,000 × *g* for 1 min, the supernatant was removed. The cell pellet was stored in a -20°C freezer until DNA extraction. For RNA extraction, triplicate 0.3-ml aliquots were transferred to 1.5-ml DNase- and RNase-free tubes, and each aliquot was mixed vigorously with 0.6 ml of RNeasy Protect Bacteria reagent (Qiagen, Hilden, Germany). After centrifugation at 5,000 × *g* for 10 min, the supernatant was carefully removed and the pellet was stored at -80°C until further treatment.

DNA/RNA extraction and purification procedure. DNA extraction from the pelleted cell culture samples were performed with the DNeasy blood and tissue kit (Qiagen) according to the protocol provided by the manufacturer. The DNA samples were stored at -20°C until qPCR analyses. A two-step RT-qPCR approach was taken for analyses of *nosZ* transcription according to the established protocol (62). Before extraction, 1 μl of luciferase control mRNA (Promega, Madison, WI) diluted to 10¹⁰ copies · μl⁻¹ was added to each sample as an internal standard to account for RNA loss during extraction, purification, and reverse transcription processes (63). The recovery of the control RNA was used to check for the validity of transcription analyses. A mixture of 350 μl of RLT buffer from an RNeasy minikit (Qiagen) and 7 μl of β-mercaptoethanol was added to the tubes containing cell pellets, and the tubes were vortexed for 10 s. Each suspension was transferred to a 2.0-ml reinforced tube containing 0.1-mm-diameter glass beads (Omni International, Kennesaw, GA). Cells were disrupted for 5 min at 5 m/s in a Bead Ruptor 12 homogenizer (Omni International). Total RNA was extracted using an RNeasy minikit according to the protocol provided by the manufacturer. The eluents were then treated with RNase-free DNase I (Qiagen) to remove residual DNA (64). The DNase-treated samples were purified using the RNeasy MinElute cleanup kit (Qiagen), and 20-μl eluent volumes were collected after purification. To 10 μl of the eluents, 1 μl of a 10 mM dinucleoside triphosphate (dNTP) mixture (Invitrogen, Waltham, MA) and 2 μl of random hexamers (Invitrogen) were added. The remaining eluents were stored at -20°C and later employed to check for DNA contamination by using qPCR assays targeting the *nosZ* gene. The mixture was incubated at 65°C for 5 min. After cooling in ice for 1 min, 4 μl of 5× First-Strand buffer, 1 μl of 0.1 M dithiothreitol (DTT; Invitrogen), and 1 μl of RNaseOUT solution (Invitrogen) were added, and the reaction mixture was incubated at room temperature for 2 min. After 1 μl of SuperScript III reverse transcriptase (Invitrogen) was added, the reaction mixture was incubated with the following temperature cycle: 10 min at room temperature, 3 h at 42°C, and 15 min at 70°C. The resulting cDNA was chilled on ice before 1 μl of RNase H (Invitrogen) was added to digest RNA strands of RNA-cDNA hybrids and residual RNA in the solution.

TABLE 1 Primers used for qPCR and RT-qPCR analyses and their qPCR calibration curve parameters

Primer	Primer sequence (5' to 3')	Target gene (locus tag)	Amplicon length (bp)	Slope	y intercept	Amplification efficiency	R ²	Reference or source
<i>GenosZf1000</i>	TCGATCTACTTCTGCCGAC	<i>nosZ</i> (GAU_1385)	151	-3.523	36.21	92.2	0.993	This study
<i>GenosZr1150</i>	CGAACGCCTGATCCTTGATG							
<i>GerecAf780</i>	CGACATCATGTACGCGGAAG	<i>recA</i> (GAU_1917)	183	-3.449	35.99	95.0	0.999	This study
<i>GerecAr972</i>	CTTCACCTGTCTCGACCT							
<i>lucf</i>	TACAACACCCCAACATCTTCGA	Luciferase control mRNA	67	-3.32	34.69	100.3	1	62
<i>lucr</i>	GGAAGTTCACCGGCGTCAT							

The reaction mixture was incubated at 37°C for 20 min, and the resulting cDNA solution was diluted 5-fold with the addition of 84 μ l of nuclease-free water to reduce inhibitory effects of the reagents on qPCR. The triplicate samples collected at each time point were treated independently through extraction, purification, and reverse transcription procedures and yielded three separate cDNA samples later subjected to qPCR.

Quantitative PCR/RT-qPCR assays for cell number determination and analyses of *nosZ* transcription. Quantification of *nosZ* genes in the genomic DNA and the cDNA samples were performed with quantitative PCR using established protocols, with modifications (62) (Fig. 5). *recA*, the housekeeping gene encoding recombinase A, was quantified in the cDNA samples for normalization of the RT-qPCR data, and the *nosZ* gene expression levels were presented as the *nosZ* transcript copy number per *recA* transcript. The new primers used in this study were designed using the Primer3 software, and amplicon sizes were limited to <200 bp for accurate quantification (Table 1) (65). SYBR green detection chemistry was used for qPCR assays performed with the QuantStudio 3 real-time PCR instrument (Thermo Fisher Scientific, Waltham, MA). The reaction mixture was prepared with 2 \times Power SYBR green PCR master mix solution (Applied Biosystems, Waltham, MA). The calibration curves were constructed with a serial dilution series of the target fragments inserted into PCR2.1 vectors using the TOPO TA cloning kit (Thermo Fisher Scientific). The qPCR assay for all three target genes yielded consistent results for DNA copy numbers as low as 10¹ copies/ μ l. Amplification of the no-template controls and DNase-treated RNA samples yielded negative results. Consistent melting curves indicated the target specificity of the qPCRs. The RNA recovery rates, as determined from the recovery of the *luc* control, ranged from 19.61% to 41.28%. For each DNA or cDNA sample analyzed, qPCRs were performed in triplicate, and the average threshold cycle (C_T) values of these technical replicates were obtained for calculation of the copy number.

Statistical analyses. Statistical analyses (*t* tests) were performed with SPSS 24 software (IBM Corp., NY, USA). Unless otherwise mentioned, the presented data are the averages and standard deviations of the results from triplicate experiments. *P* values of <0.05 were considered significant.

SUPPLEMENTAL MATERIAL

Supplemental material for this article may be found at <https://doi.org/10.1128/AEM.00502-17>.

SUPPLEMENTAL FILE 1, PDF file, 0.1 MB.

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REFERENCES

- Montzka SA, Dlugokencky EJ, Butler JH. 2011. Non-CO₂ greenhouse gases and climate change. *Nature* 476:43–50. <https://doi.org/10.1038/nature10322>.
- Lashof DA, Ahuja DR. 1990. Relative contributions of greenhouse gas emissions to global warming. *Nature* 344:529–531. <https://doi.org/10.1038/344529a0>.
- Forster P, Ramaswamy V, Artaxo P, Bernsten T, Betts R, Fahey DW, Haywood J, Lean J, Lowe DC, Myhre G, Nganga J, Prinn R, Raga G, Schultz M, van Dorland R. 2007. Changes in atmospheric constituents and in radiative forcing, p 129–234. In Solomon S, Qin D, Manning M, Chen Z, Marquis M, Averyt KB, Tingor M, Miller HL (ed), *Climate change 2007: the physical science basis. Contribution of working group I to the fourth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, United Kingdom.
- Intergovernmental Panel on Climate Change. 2014. *Climate change 2014: mitigation of climate change. Contribution of working group III to the fifth assessment report of the Intergovernmental Panel on Climate Change*. Cambridge University Press, Cambridge, United Kingdom, and New York, NY.
- Ravishankara AR, Daniel JS, Portmann RW. 2009. Nitrous oxide (N₂O): the dominant ozone-depleting substance emitted in the 21st century. *Science* 326:123–125. <https://doi.org/10.1126/science.1176985>.
- Portmann RW, Daniel JS, Ravishankara AR. 2012. Stratospheric ozone depletion due to nitrous oxide: influences of other gases. *Philos Trans R Soc Lond B Biol Sci* 367:1256–1264. <https://doi.org/10.1098/rstb.2011.0377>.
- Butterbach-Bahl K, Baggs EM, Dannenmann M, Kiese R, Zechmeister-Boltenstern S. 2013. Nitrous oxide emissions from soils: how well do we understand the processes and their controls? *Philos Trans R Soc Lond B Biol Sci* 368:20130122. <https://doi.org/10.1098/rstb.2013.0122>.
- Thomson AJ, Giannopoulos G, Pretty J, Baggs EM, Richardson DJ. 2012. Biological sources and sinks of nitrous oxide and strategies to mitigate emissions. *Philos Trans R Soc Lond B Biol Sci* 367:1157–1168. <https://doi.org/10.1098/rstb.2011.0415>.
- Zhu X, Burger M, Doane TA, Horwath WR. 2013. Ammonia oxidation pathways and nitrifier denitrification are significant sources of N₂O and NO under low oxygen availability. *Proc Natl Acad Sci U S A* 110: 6328–6333. <https://doi.org/10.1073/pnas.1219993110>.
- Bateman EJ, Baggs EM. 2005. Contributions of nitrification and denitri-

- fication to N₂O emissions from soils at different water-filled pore space. *Biol Fertil Soils* 41:379–388. <https://doi.org/10.1007/s00374-005-0858-3>.
11. Philippot L, Andert J, Jones CM, Bru D, Hallin S. 2011. Importance of denitrifiers lacking the genes encoding the nitrous oxide reductase for N₂O emissions from soil. *Glob Chang Biol* 17:1497–1504. <https://doi.org/10.1111/j.1365-2486.2010.02334.x>.
 12. Cruz-García C, Murray AE, Klappenbach JA, Stewart V, Tiedje JM. 2007. Respiratory nitrate ammonification by *Shewanella oneidensis* MR-1. *J Bacteriol* 189:656–662. <https://doi.org/10.1128/JB.01194-06>.
 13. Jones LC, Peters B, Lezama Pacheco JS, Casciotti KL, Fendorf S. 2015. Stable isotopes and iron oxide mineral products as markers of chemo-denitrification. *Environ Sci Technol* 49:3444–3452. <https://doi.org/10.1021/es504862x>.
 14. Mania D, Heylen K, van Spanning RJ, Frostegard A. 2014. The nitrate-ammonifying and *nosZ* carrying bacterium *Bacillus vireti* is a potent source and sink for nitric and nitrous oxides under high nitrate conditions. *Environ Microbiol* 16:3196–3210. <https://doi.org/10.1111/1462-2920.12478>.
 15. Sanford RA, Wagner DD, Wu Q, Chee-Sanford JC, Thomas SH, Cruz-García C, Rodríguez G, Massol-Deyá A, Krishnani KK, Ritalahti KM, Nissen S, Konstantinidis KT, Löffler FE. 2012. Unexpected nondenitrifier nitrous oxide reductase gene diversity and abundance in soils. *Proc Natl Acad Sci U S A* 109:19709–19714. <https://doi.org/10.1073/pnas.1211238109>.
 16. Jones CM, Spor A, Brennan FP, Breuil M-C, Bru D, Lemanceau P, Griffiths B, Hallin S, Philippot L. 2014. Recently identified microbial guild mediates soil N₂O sink capacity. *Nat Clim Chang* 4:801–805. <https://doi.org/10.1038/nclimate2301>.
 17. Yoon S, Nissen S, Park D, Sanford RA, Löffler FE. 2016. Nitrous oxide reduction kinetics distinguish bacteria harboring clade I versus clade II *NosZ*. *Appl Environ Microbiol* 82:3793–3800. <https://doi.org/10.1128/AEM.00409-16>.
 18. Jones CM, Graf DRH, Bru D, Philippot L, Hallin S. 2013. The unaccounted yet abundant nitrous oxide-reducing microbial community: a potential nitrous oxide sink. *ISME J* 7:417–426. <https://doi.org/10.1038/ismej.2012.125>.
 19. Graf DR, Jones CM, Hallin S. 2014. Intergenic comparisons highlight modularity of the denitrification pathway and underpin the importance of community structure for N₂O emissions. *PLoS One* 9:e114118. <https://doi.org/10.1371/journal.pone.0114118>.
 20. Domeignoz-Horta L, Spor A, Bru D, Breuil M-C, Bizouard F, Leonard J, Philippot L. 2015. The diversity of the N₂O reducers matters for the N₂O:N₂ denitrification end-product ratio across an annual and a perennial cropping system. *Front Microbiol* 6:971. <https://doi.org/10.3389/fmicb.2015.00971>.
 21. Domeignoz-Horta LA, Putz M, Spor A, Bru D, Breuil MC, Hallin S, Philippot L. 2016. Non-denitrifying nitrous oxide-reducing bacteria—an effective N₂O sink in soil. *Soil Biol Biochem* 103:376–379. <https://doi.org/10.1016/j.soilbio.2016.09.010>.
 22. Orellana LH, Rodríguez-R LM, Higgins S, Chee-Sanford JC, Sanford RA, Ritalahti KM, Löffler FE, Konstantinidis KT. 2014. Detecting nitrous oxide reductase (*nosZ*) genes in soil metagenomes: method development and implications for the nitrogen cycle. *mBio* 5(3):e01193-14. <https://doi.org/10.1128/mBio.01193-14>.
 23. de Bruin WP, Kotterman MJ, Posthumus MA, Schraa G, Zehnder AJ. 1992. Complete biological reductive transformation of tetrachloroethene to ethane. *Appl Environ Microbiol* 58:1996–2000.
 24. Shen Z, Wang D, Ruan Y, Xue C, Zhang J, Li R, Shen Q. 2014. Deep 16S rRNA pyrosequencing reveals a bacterial community associated with banana fusarium wilt disease suppression induced by bio-organic fertilizer application. *PLoS One* 9:e98420. <https://doi.org/10.1371/journal.pone.0098420>.
 25. Zeng Y, Selyanin V, Lukeš M, Dean J, Kaftan D, Feng F, Koblížek M. 2015. Characterization of the microaerophilic, bacteriochlorophyll a-containing bacterium *Gemmatimonas phototrophica* sp. nov., and emended descriptions of the genus *Gemmatimonas* and *Gemmatimonas aurantiaca*. *Int J Syst Evol Microbiol* 65:2410–2419. <https://doi.org/10.1099/ij.s.0.000272>.
 26. Zhang H, Sekiguchi Y, Hanada S, Hugenholtz P, Kim H, Kamagata Y, Nakamura K. 2003. *Gemmatimonas aurantiaca* gen. nov., sp. nov., a Gram-negative, aerobic, polyphosphate-accumulating micro-organism, the first cultured representative of the new bacterial phylum *Gemmatimonadetes* phyl. nov. *Int J Syst Evol Microbiol* 53:1155–1163. <https://doi.org/10.1099/ij.s.0.02520-0>.
 27. DeBruyn JM, Fawaz MN, Peacock AD, Dunlap JR, Nixon LT, Cooper KE, Radosevich M. 2013. *Gemmatirosa kalamazooensis* gen. nov., sp. nov., a member of the rarely-cultivated bacterial phylum *Gemmatimonadetes*. *J Gen Appl Microbiol* 59:305–312. <https://doi.org/10.2323/jgama.59.305>.
 28. Davis KER, Joseph SJ, Janssen PH. 2005. Effects of growth medium, inoculum size, and incubation time on culturability and isolation of soil bacteria. *Appl Environ Microbiol* 71:826–834. <https://doi.org/10.1128/AEM.71.2.826-834.2005>.
 29. Janssen P. 2006. Identifying the dominant soil bacterial taxa in libraries of 16S rRNA and 16S rRNA genes. *Appl Environ Microbiol* 72:1719–1728. <https://doi.org/10.1128/AEM.72.3.1719-1728.2006>.
 30. DeBruyn JM, Nixon LT, Fawaz MN, Johnson AM, Radosevich M. 2011. Global biogeography and quantitative seasonal dynamics of *Gemmatimonadetes* in soil. *Appl Environ Microbiol* 77:6295–6300. <https://doi.org/10.1128/AEM.05005-11>.
 31. Takaichi S, Maoka T, Takasaki K, Hanada S. 2010. Carotenoids of *Gemmatimonas aurantiaca* (*Gemmatimonadetes*): identification of a novel carotenoid, deoxyoscillo 2-rhamnoside, and proposed biosynthetic pathway of oscillo 2,2'-dirhamnoside. *Microbiology* 156:757–763. <https://doi.org/10.1099/mic.0.034249-0>.
 32. Zeng Y, Feng F, Medová H, Dean J, Koblížek M. 2014. Functional type 2 photosynthetic reaction centers found in the rare bacterial phylum *Gemmatimonadetes*. *Proc Natl Acad Sci U S A* 111:7795–7800. <https://doi.org/10.1073/pnas.1400295111>.
 33. Hu H-W, Chen D, He J-Z. 2015. Microbial regulation of terrestrial nitrous oxide formation: understanding the biological pathways for prediction of emission rates. *FEMS Microbiol Rev* 39:729–749. <https://doi.org/10.1093/femsre/fuv021>.
 34. Geets J, Boon N, Verstraete W. 2006. Strategies of aerobic ammonia-oxidizing bacteria for coping with nutrient and oxygen fluctuations. *FEMS Microbiol Ecol* 58:1–13. <https://doi.org/10.1111/j.1574-6941.2006.00170.x>.
 35. Kartal B, Wessels HJCT, van der Biezen E, Francoijs K-J, Jetten MSM, Klotz MG, Stein LY. 2012. Effects of nitrogen dioxide and anoxia on global gene and protein expression in long-term continuous cultures of *Nitrosomonas eutropha* C91. *Appl Environ Microbiol* 78:4788–4794. <https://doi.org/10.1128/AEM.00668-12>.
 36. Fischer M, Falke D, Pawlik T, Sawers RG. 2014. Oxygen-dependent control of respiratory nitrate reduction in mycelium of *Streptomyces coelicolor* A3(2). *J Bacteriol* 196:4152–4162. <https://doi.org/10.1128/JB.02202-14>.
 37. van Keulen G, Alderson J, White J, Sawers RG. 2007. The obligate aerobic actinomycete *Streptomyces coelicolor* A3(2) survives extended periods of anaerobic stress. *Environ Microbiol* 9:3143–3149. <https://doi.org/10.1111/j.1462-2920.2007.01433.x>.
 38. Larkin JM, Williams PM. 1978. *Runella slithyformis* gen. nov., sp. nov., a curved, nonflexible, pink bacterium. *Int J Syst Evol Microbiol* 28:32–36.
 39. Daligault H, Lapidus A, Zeytun A, Nolan M, Lucas S, Del Rio TG, Tice H, Cheng J-F, Tapia R, Han C, Goodwin L, Pitluck S, Liolios K, Pagani I, Ivanova N, Huntemann M, Mavromatis K, Mikhailova N, Pati A, Chen A, Palaniappan K, Land M, Hauser L, Brambilla E-M, Rohde M, Verberg S, Göker M, Bristow J, Eisen JA, Markowitz V, Hugenholtz P, Kyrpides NC, Klenk H-P, Woyske T. 2011. Complete genome sequence of *Halicomonobacter hydrossis* type strain (O^T). *Stand Genomic Sci* 4:352–360. <https://doi.org/10.4056/signs.1964579>.
 40. Pagani I, Chertkov O, Lapidus A, Lucas S, Glavina Del Rio T, Tice H, Copeland A, Cheng J-F, Nolan M, Saunders E, Pitluck S, Held B, Goodwin L, Liolios K, Ovchinnikova G, Ivanova N, Mavromatis K, Pati A, Chen A, Palaniappan K, Land M, Hauser L, Jeffries CD, Detter JC, Han C, Tapia R, Ngatchou-Djao OD, Rohde M, Göker M, Spring S, Sikorski J, Woyske T, Bristow J, Eisen JA, Markowitz V, Hugenholtz P, Klenk H-P, Kyrpides NC. 2011. Complete genome sequence of *Mariivirga tractuosa* type strain (H-43^T). *Stand Genomic Sci* 4:154–162. <https://doi.org/10.4056/signs.1623941>.
 41. Perolat P, Chappel RJ, Adler B, Baranton G, Bulach DM, Billingham ML, Letocart M, Merien F, Serrano MS. 1998. *Leptospira fainei* sp. nov., isolated from pigs in Australia. *Int J Syst Evol Microbiol* 48:851–858.
 42. Zumft WG. 1997. Cell biology and molecular basis of denitrification. *Microbiol Mol Biol Rev* 61:533–616.
 43. Zumft WG, Kroneck PM. 2006. Respiratory transformation of nitrous oxide (N₂O) to dinitrogen by bacteria and archaea, p 107–227. *In* Robert KP (ed), *Advances in microbiology physiology*, vol 52. Academic Press, Cambridge MA.
 44. Payne WJ, Grant MA, Shapleigh J, Hoffman P. 1982. Nitrogen oxide

- reduction in *Wolinella succinogenes* and *Campylobacter* species. *J Bacteriol* 152:915–918.
45. Sellers MJ, Hall SJ, Kelly DJ. 2002. Growth of *Campylobacter jejuni* supported by respiration of fumarate, nitrate, nitrite, trimethylamine-*N*-oxide, or dimethyl sulfoxide requires oxygen. *J Bacteriol* 184:4187–4196. <https://doi.org/10.1128/JB.184.15.4187-4196.2002>.
 46. Layer G, Pierik AJ, Trost M, Rigby SE, Leech HK, Grage K, Breckau D, Astner I, Jänsch L, Heathcote P, Warren MJ, Heinz DW, Jahn D. 2006. The substrate radical of *Escherichia coli* oxygen-independent coproporphyrinogen III oxidase HemN. *J Biol Chem* 281:15727–15734. <https://doi.org/10.1074/jbc.M512628200>.
 47. Betlach MR, Tiedje JM. 1981. Kinetic explanation for accumulation of nitrite, nitric oxide, and nitrous oxide during bacterial denitrification. *Appl Environ Microbiol* 42:1074–1084.
 48. Morley N, Baggs EM, Dörsch P, Bakken L. 2008. Production of NO, N₂O and N₂ by extracted soil bacteria, regulation by NO₂⁻ and O₂ concentrations. *FEMS Microbiol Ecol* 65:102–112. <https://doi.org/10.1111/j.1574-6941.2008.00495.x>.
 49. Torres MJ, Rubia MI, Bedmar EJ, Delgado MJ. 2011. Denitrification in *Sinorhizobium meliloti*. *Biochem Soc Trans* 39:1886–1889. <https://doi.org/10.1042/BST20110733>.
 50. Bergaust L, Mao Y, Bakken L, Frostegård A. 2010. Denitrification response patterns during the transition to anoxic respiration and posttranscriptional effects of suboptimal pH on nitrous oxide reductase in *Paracoccus denitrificans*. *Appl Environ Microbiol* 76:6387–6396. <https://doi.org/10.1128/AEM.00608-10>.
 51. Torres M, Rubia M, de la Pena T, Pueyo J, Bedmar E, Delgado M. 2014. Genetic basis for denitrification in *Ensifer meliloti*. *BMC Microbiol* 14:142. <https://doi.org/10.1186/1471-2180-14-142>.
 52. Qu Z, Bakken LR, Molstad L, Frostegård Å, Bergaust LL. 2016. Transcriptional and metabolic regulation of denitrification in *Paracoccus denitrificans* allows low but significant activity of nitrous oxide reductase under oxic conditions. *Environ Microbiol* 18:2951–2963. <https://doi.org/10.1111/1462-2920.13128>.
 53. Davidson EA, Verchot LV, Cattânio JH, Ackerman IL, Carvalho J. 2000. Effects of soil water content on soil respiration in forests and cattle pastures of eastern Amazonia. *Biogeochemistry* 48:53–69. <https://doi.org/10.1023/A:1006204113917>.
 54. Glatzel S, Stahr K. 2001. Methane and nitrous oxide exchange in differently fertilized grassland in southern Germany. *Plant Soil* 231:21–35. <https://doi.org/10.1023/A:1010315416866>.
 55. Law Y, Ye L, Pan Y, Yuan Z. 2012. Nitrous oxide emissions from wastewater treatment processes. *Philos Trans R Soc B* 367:1265–1277. <https://doi.org/10.1098/rstb.2011.0317>.
 56. Bollmann A, Conrad R. 1998. Influence of O₂ availability on NO and N₂O release by nitrification and denitrification in soils. *Glob Chang Biol* 4:387–396. <https://doi.org/10.1046/j.1365-2486.1998.00161.x>.
 57. Qu Z, Wang J, Almøy T, Bakken LR. 2014. Excessive use of nitrogen in Chinese agriculture results in high N₂O/(N₂O + N₂) product ratio of denitrification, primarily due to acidification of the soils. *Glob Chang Biol* 20:1685–1698. <https://doi.org/10.1111/gcb.12461>.
 58. Granger J, Ward BB. 2003. Accumulation of nitrogen oxides in copper-limited cultures of denitrifying bacteria. *Limnol Oceanogr* 48:313–318. <https://doi.org/10.4319/lo.2003.48.1.0313>.
 59. Wolin E, Wolin MJ, Wolfe R. 1963. Formation of methane by bacterial extracts. *J Biol Chem* 238:2882–2886.
 60. Yoon S, Sanford RA, Löffler FE. 2013. *Shewanella* spp. use acetate as an electron donor for denitrification but not ferric iron or fumarate reduction. *Appl Environ Microbiol* 79:2818–2822. <https://doi.org/10.1128/AEM.03872-12>.
 61. Sander R. 2015. Compilation of Henry's law constants (version 4.0) for water as solvent. *Atmos Chem Phys* 15:4399–4981. <https://doi.org/10.5194/acp-15-4399-2015>.
 62. Yoon S, Cruz-García C, Sanford R, Ritalahti KM, Löffler FE. 2015. Denitrification versus respiratory ammonification: environmental controls of two competing dissimilatory NO₃⁻/NO₂⁻ reduction pathways in *Shewanella loihica* strain PV-4. *ISME J* 9:1093–1104. <https://doi.org/10.1038/ismej.2014.201>.
 63. Amos BK, Ritalahti KM, Cruz-García C, Padilla-Crespo E, Löffler FE. 2008. Oxygen effect on *Dehalococcoides* viability and biomarker quantification. *Environ Sci Technol* 42:5718–5726. <https://doi.org/10.1021/es703227g>.
 64. Johnson DR, Lee PK, Holmes VF, Alvarez-Cohen L. 2005. An internal reference technique for accurately quantifying specific mRNAs by real-time PCR with application to the *tceA* reductive dehalogenase gene. *Appl Environ Microbiol* 71:3866–3871. <https://doi.org/10.1128/AEM.71.7.3866-3871.2005>.
 65. Untergasser A, Cutcutache I, Koressaar T, Ye J, Faircloth BC, Remm M, Rozen SG. 2012. Primer3—new capabilities and interfaces. *Nucleic Acids Res* 40:e115. <https://doi.org/10.1093/nar/gks596>.